

# Nativeness as gradient: Towards a more complete value assessment of species in a rapidly changing world

Avery P. Hill<sup>1,2\*</sup>

August 13, 2024

<sup>1</sup> Center for Biodiversity and Community Science, California Academy of Sciences, 55 Music Concourse Drive, San Francisco, 94118, CA, U.S.A.

<sup>2</sup> Department of Biology, Stanford University, 371 Jane Stanford Way, Stanford, 94305, CA, U.S.A.

Corresponding author(s). E-mail(s): [ahill@calacademy.org](mailto:ahill@calacademy.org);

## Acknowledgments

A grant from the Gordon and Betty Moore Foundation (ID: GBMF8449) facilitated completion of this work. I am immensely grateful to all that provided critical feedback and insight during the development of this manuscript, including Gili Greenbaum, Christopher Field, Elizabeth Hadly, Tadashi Fukami, Lucas Pavan, Tatiana Bellagio, Maria Viteri, Emily Russell, and Nona Chiariello. Additional thanks to Pamela Thornell for providing the strong foundation from which I conducted this work.

# Nativeness as gradient: Towards a more complete value assessment of species in a rapidly changing world

## Abstract

1 Conservation biologists recognize a duty to maintain as much value as possible in ecosystems that are  
2 threatened by recent anthropogenic impacts. Until recently the paradigm of contemporary conservation  
3 seemed relatively straightforward: the best way to maintain the value of species and ecosystems  
4 at a given location was to maintain—or shepherd the system back towards—historical conditions.  
5 Among the most difficult theoretical tasks was the determination of “baseline” historical conditions (or  
6 trajectories) to return to, recognizing the dynamism of ecosystems over time. However, the rate, scale,  
7 and magnitude of contemporary climate change, species introductions, and land-use change make it  
8 increasingly impractical to return locations to any kind of historical state. This forces a paradigm shift  
9 which is both ongoing and difficult, and necessitates a rigorous evaluation of the scientific and ethical  
10 foundations of modern conservation along with a careful reexamination of terminology. Here, I discuss  
11 the moral relevance and waning utility of the geographically-based and dichotomous understanding of  
12 “native” (or “*in situ*”) which is an important component of conservation ethics and practice. I then  
13 propose a new understanding of nativeness in which a species is native—not to a geographic location—  
14 but to a quantifiable set of biotic, climatic, geologic, and topographic conditions (i.e. its niche) that  
15 can then map to geographic space. Following this, I demonstrate the unique utility of this concept,  
16 which I will refer to as “econativeness,” in thinking through conservation problems—range expansions,  
17 range contractions, species introductions, and assisted migration—where the classical understanding  
18 of nativeness has become increasingly inadequate for assessing the moral value of species.

## 19 1 Introduction

20 Maintaining the value of species and ecosystems (or at least mitigating the loss of it) is the moral foun-  
21 dation of conservation projects [Soule, 1985]. There are varied and sometimes competing approaches  
22 which emphasize different values and suggest different conservation actions. One framework empha-  
23 sizes the values of ‘naturalness’ and suggests actions that primarily mitigate human impact. Another  
24 emphasizes the myriad values of biodiversity and favors actions that primarily increase or maintain  
25 biodiversity. Yet another emphasizes values of ecological and evolutionary situatedness and primarily  
26 encourages the protection of historical ecological structure and processes. These frameworks rely on  
27 a number of shared values (differing in emphasis) and many of these values are at least partly depen-  
28 dent on a species continued existence in the ecosystem to which it is adapted (i.e. its nativeness).  
29 This dependence on nativeness is part of the reason conservationists don’t think of zoos as an end-  
30 goal—keeping the species alive may preserve some value, but much is lost with the loss of ecological,  
31 evolutionary, and environmental relationships. As a result, most conservation projects work towards  
32 maintaining an ecosystem’s species, their relationships, and therefore the greatest amount of value in  
33 the face of anthropogenic impacts that might erode them.

34 This has primarily been conducted in an historically-oriented and place-based manner: at a given  
35 location, the most value is thought to be maintained if unaffected areas remain unaffected and affected  
36 areas are shepherded back toward the historical conditions of the location. This practice makes sense  
37 when the following conditions are met: (1) the species value is maximized when in the environment  
38 in which it is adapted, (2) the described ‘historical conditions’ or ‘baseline’ of a site are those that  
39 the species is best adapted to, and finally (3) a site can be managed towards the described historical  
40 conditions.

41 Condition 1 is commonly understood to hold true—think again of how zoos and botanical gardens  
42 fall short of central conservation goals. Condition 2, concerning the description of baseline conditions  
43 of a site, is the subject of much concern in conservation and ecology [Jachowski et al., 2015]. When  
44 this condition is violated, the conservation paradigm can fault. For example, a misunderstanding of  
45 the ecological and evolutionary forces that shaped the giant sequoia tree (*Sequoiadendron giganteum*)  
46 contributed to fire suppression which directly hindered the germination of giant sequoia seedlings for  
47 nearly a century [U.S. National Park Service, 2021]. Misunderstandings like these become less frequent  
48 as biologists learn more about the drivers of ecosystem structure and function, and Condition 2 can

49 more often be met. Condition 3, the ability to maintain or manage a site towards historical conditions,  
50 may follow the opposite trajectory and become more difficult to satisfy as time passes. And perhaps this  
51 is the greatest threat to the traditional conservation paradigm: the rapidly diminishing likelihood that  
52 ecosystems can be returned to any kind of historical condition, due to a combination of anthropogenic  
53 drivers including climate change, land-use change, and species introductions [Camacho et al., 2010,  
54 Camacho, 2010b].

55 Global ecosystems may be approaching a tipping point in response to contemporary human activity  
56 [Barnosky et al., 2012]. Land cover types of human origin (e.g. agricultural land and cities) now  
57 blanket more than half of the earth’s land surface [Hooke and Martín-Duque, 2012], and the rapidly  
58 increasing rate of species introductions has resulted in the establishment of more than 37,000 species  
59 outside of their historical ranges [Roy et al., 2023]. In some cases, these impacts may be reversible.

60 The impacts of climate change, on the other hand, will be patently irreversible in many ecosystems.  
61 Since 1850, industrial human activity has added 2,390 gigatons of CO<sub>2</sub> to the atmosphere, raising the  
62 atmospheric concentration of CO<sub>2</sub> from 280 to 410 parts per million. Consequently, global mean  
63 surface temperatures have increased by more than 1°C, arctic sea ice in September has decreased in  
64 area by 40%, and drought events have increased in severity and frequency worldwide.

65 These impacts are only expected to increase in both scope and severity as climate change con-  
66 tinues, and the extent to which the impacts of climate change may be considered reversible is de-  
67 pendent on what we think of as being “reversed.” The erosion of general ecological stability may be  
68 reversible, but the eco-evolutionary processes set in motion by global shifts in temperature, precipi-  
69 tation, and disturbance regimes—evidenced by evolutionary responses and shifts in the distributions  
70 of myriad species—are in many ways not. The body of work documenting observed and expected  
71 regime shifts in ecosystems is diverse, expansive, and growing [Walker et al., 2023, Scheffer et al., 2009,  
72 Reyer et al., 2015, Barnosky et al., 2012]. Altogether, the state of global environmental change and its  
73 impact on ecosystems presents a significant challenge to place-based and history-oriented conservation  
74 practices—specifically those that are motivated by the nativeness of a species.

75 Terms like “*in situ*,” and “native habitat” have almost always been used in a geographically de-  
76 pendent and dichotomous way. Species are either native or non-native, and localized species databases  
77 like California’s plant occurrence database Calflora are rigorous in their dichotomous classification of  
78 various plants [Calflora, 2022]. Conservationists use this native/non-native dichotomy because it has  
79 been useful in our understanding of species invasiveness and their effect on ecological stability. While  
80 this dichotomy has been useful, its utility becomes increasingly eroded due to the velocity of global  
81 environmental change. How can “native” species sometimes behave as invasive within their historical  
82 range [Nackley et al., 2017], like the explosive population growth of both the mountain pine beetle  
83 (*Dendroctonus spp.*) [Bentz et al., 2010] and Australian sandplain heath (*Allocasuarina huegeliana*)  
84 [Shackelford et al., 2013]? Are species that migrate in response to climate change non-native in their  
85 new locations and do they lose all the value that their historical range confers? And what does “native”  
86 mean for a narrow-ranged endemic species when it is unable to survive in its historical habitat due to  
87 climate change?

88 Recent work has highlighted the shortcomings of terms like “native” habitat for describing a species  
89 eco-evolutionary situatedness in a rapidly changing world. Some have cited the inability of “native”  
90 to properly account for species that shift their distribution in response to climate change, and sug-  
91 gest adding a third class to the native/non-native dichotomy, “neonative,” to describe these species  
92 [Essl et al., 2019]. This contribution moves us away from a strict dichotomy, but its lack of gradient  
93 keeps it susceptible to classification problems [Wilson, 2020]. Furthermore, no attempt has been made  
94 to describe the difference in value between native, neonative, and non-native species, and it’s unclear  
95 how this could be accomplished.

96 Others have criticized the temporal ambiguity of nativeness (e.g. the species present in the Americas  
97 and Australia before 1492 and 1770, respectively, are considered native, but the correct date for  
98 somewhere like Europe remains unclear) and spatial ambiguity (e.g. it’s difficult to detect range edges  
99 and some species have inherently patchy distributions) [Hill and Hadly, 2018, Warren, 2007]. These  
100 criticisms have inspired gestures towards re-framing the “native” concept to include the niche and  
101 the dynamism of environment [Pereyra, 2020] and more rigorous efforts to classify different states of  
102 nativeness from a paleontological perspective [Crees and Turvey, 2015].

103 I build on this and on previous work which suggests that “native” is increasingly unhelpful while it is  
104 both dichotomous and geographically-based [Hill and Hadly, 2018, Pereyra, 2020], and argue that the

105 eco-evolutionary situatedness of a species—and the value that is conferred thereby—is best evaluated  
106 by a niche-based, nonspatial, gradated conception of “nativeness” which can then be subsequently  
107 translated to geographic space. The novel conception of nativeness I offer below, called “econative,”  
108 aims to reconcile our traditional understanding of “native” with its modern criticisms and alternatives  
109 to produce a more coherent, unified, and useful definition to describe where species belong in a rapidly  
110 changing world.

## 111 2 Introducing “Econative”

112 I propose that a species is *not* native to a geographic location, rather, it is native to a quantifiable  
113 set of biotic, climatic, geologic, and topographic conditions (i.e. its niche). Furthermore, a species  
114 proximity to its native niche is a gradient rather than a dichotomy. It’s easy to see why “native” has  
115 come to refer to particular geographic regions: at short timescales and minimal environmental change,  
116 the environmental conditions to which a species is adapted is geographically stable. However, over  
117 longer time or during periods of rapid change the relationship between geography and a specific set of  
118 environmental conditions decouples. For example, we generally consider giant sequoias (*Sequoiadendron*  
119 *giganteum*) native to the western slope of the Sierra Nevada mountains of California, but millions of  
120 years ago these trees were probably distributed across a much larger region in North America, perhaps  
121 as far as the modern state of Idaho [Lowe, 2014]. What then, is the native distribution of these  
122 trees? If a geographic area had to be delineated using the classical understanding of nativeness, a  
123 biogeographer would probably map the native range as the area including the current distribution and  
124 wherever else the species was recorded when Europeans arrived, whether or not the trees are found  
125 there today. Why aren’t the areas in what is now called Nevada (which had these trees around 2  
126 million years ago [Dodd and DeSilva, 2016]) included? Because the environmental conditions that the  
127 giant sequoia is adapted to are no longer present in those areas. And if the climate and fire regime  
128 changes in California over the next century such that the environmental conditions no longer support  
129 giant sequoias and the trees are entirely extirpated, will the trees still be native? An adherent to  
130 the inflexible geographically-based native/non-native dichotomy would be forced to conclude that the  
131 giant sequoias then become non-native—with repercussions for the values attributed to the species.

132 More nuanced and logically consistent is the understanding that giant sequoias are native to the  
133 environmental conditions to which they evolved in, and that these conditions intersect with geographic  
134 space over a continuum, where any given geographic location has some quantifiable measure of the  
135 native niche. Millions of years ago, the native niche of giant sequoias intersected with a larger portion  
136 of western North America, and today, the geographic instantiation of the native niche covers only a  
137 small, patchy area of the Sierra Nevada. However, we are not forced to conclude that future climate  
138 change will force giant sequoias into a “non-native” status—rather, increased temperatures and reduced  
139 snowpack will simply reduce the similarity of the current geographic distribution of giant sequoias  
140 to their native niche. Giant sequoias will effectively become “less native” than they are now, just  
141 like they are currently less native to Idaho or Nevada than they were millions of years ago. All  
142 species at all locations can be thought of as existing on a continuum between more native and less  
143 native, and this continuum can be imagined as the overlap between a species n-dimensional niche  
144 hypervolume and a location’s n-dimensional environment hypervolume (i.e. the Hutchinsonian niche  
145 concept) [Hutchinson, 1957].

146 I propose that this continuum between more native and less native, “econativeness”, is a more  
147 biologically informed and morally and ecologically useful term than “native” or “*in situ*” when referring  
148 to the location of species (Figure 1). Classical nativeness can be thought of as a special case of  
149 econativeness which applies only to the last few hundred years of plant and animal distributions—  
150 what we think of as “native” to a location today is simply what has been econative to a location for  
151 as long as modern science has been recording the distribution of species. Viewing ecological change  
152 through the lens of a human lifespan or during periods of little environmental change masks the  
153 distinction between econativeness and classical nativeness. However, the rate of recent change makes  
154 clear the incompetency of a dichotomous, place-based understanding of nativeness.

155 The econative concept depends greatly on how we understand the niche, and the details have  
156 important implications. The dominant ecological niche concept is derived from Hutchinson and is  
157 differentiated into the fundamental niche (everywhere a species *could* physiologically survive and re-  
158 produce in environmental space without biotic or dispersal limitations) and the realized niche (every-

159 where it *does* survive and reproduce, all factors considered) [Hutchinson, 1957]. Based solely on the  
160 Hutchinsonian niche, eonativeness is the geographic expression of both the fundamental and realized  
161 environmental hypervolumes, where the realized niche clearly maps to greater eonativeness because  
162 it describes more environmental axes of the species niche (e.g. competing species, predators, etc.).  
163 Importantly, this does not directly account for dispersal limitations, which are expressed in geographic  
164 rather than environmental space. So, if dispersal limitations and geography are not explicitly cap-  
165 tured by eonativeness, could a eucalyptus someday be similarly eonative to its Australian geography  
166 and its introduced California range in the extremely unlikely event that all co-occurring species and  
167 environmental components are similarly present in both locations?

168 No, not if we consider burgeoning developments in the niche concept, which move towards more  
169 completely capturing the eco-evolutionary situatedness of a species. A key task in ‘bringing the  
170 Hutchinsonian niche into the 21st century’ is describing the evolutionary component of the niche  
171 [Holt, 2009]. Recent work, especially in the field of niche conservation theory, has made important  
172 contributions to this effort [Trappes, 2021], but there is considerably more work to do until these con-  
173 cepts are neatly integrated [Morrow, 2024]. In fear of getting cut by the bleeding edge, I will gesture  
174 towards the trajectory of these ideas and their important implications for the eonative concept.

175 Trappes describes an “externalist evolutionary niche,” popular in niche construction theory, as “the  
176 (sum of the) environmental factors that lead to fitness differences in a population” [Trappes, 2021].  
177 She and others remark that this concept focuses on the external environmental forces that drive niche  
178 construction and largely ignores the (internal) forces of phenotypic change that can also shape the  
179 niche (as described in Aaby and Ramsey [Aaby and Ramsey, 2022]), and hints at an evolutionary  
180 niche concept that recognizes the niche-shaping potential of both species and environment. In my  
181 interpretation, this concept implies a complement akin to a “fundamental evolutionary niche” that  
182 corresponds to a hypervolume in n-dimensional environmental space to which a species *could* have  
183 have become adapted in its eco-evolutionary history. This suggests a kind of buffer zone in environ-  
184 mental space around the fundamental ecological niche of a species, and acknowledges the impact of  
185 geographically-adjacent environments in the development of a species ecological niche. And thus, at  
186 the frontier of niche theory, we can begin to account for geography and dispersal by considering the  
187 integration of evolution and niche construction into the ecological niche. When eonativeness refers  
188 to a niche concept like this, a eucalyptus tree could only be as eonative to California as Australia if  
189 all of the environmental conditions and co-occurring species are present *in addition to* all of the en-  
190 vironmental conditions and species the eucalyptus tree *could* have become adapted to throughout its  
191 evolutionary history. In practice, this guarantees that a species cannot be more eonative to a location  
192 beyond major biogeographic barriers (e.g. between continents) than they are to the geographic area  
193 they most recently evolved in.

194 A significant innovation provided by the eonative concept is that nativeness can begin to be quan-  
195 tified. For decades, statistical tools have been developed and refined to characterize and quantify the  
196 ecological niche. Though imperfect, the field of ecological niche modeling (ENM) has made signifi-  
197 cant strides in modeling the fundamental and realized niches of species [Elith and Leathwick, 2009].  
198 The most popular models are correlative statistical models which utilize known species occurrence  
199 data and corresponding environmental information to approximate the niche. These models output  
200 habitat suitability estimates along a continuous gradient, and naturally correspond to many of the  
201 abiotic dimensions of eonativeness. However, the fundamental niche that most ENMs attempt to  
202 capture is incomplete, and is commonly limited to a handful of bioclimatic variables. Recent work  
203 has acknowledged the importance of—and made progress towards—including evolutionary processes  
204 [Bush et al., 2016] and biotic interactions [Wisz et al., 2013] in ENMs. As the field progresses, we  
205 will be able to more completely model a species niche, express that niche onto geographic space, and  
206 calculate a more complete eonativeness score. The current capabilities of ENM provide a strong  
207 foundation—and for the practical application of eonativeness we might proxy the biotic and evo-  
208 lutionary dimensions of the niche by both adding environmental buffer space around the modeled  
209 fundamental niche and incorporating local species composition.

210 The result of applying ENM to the calculation of eonativeness would be an eonativeness score on  
211 a gradient between 0 and 1, similar to the habitat suitability scores usually output by ENMs. Though  
212 the eonativeness of a species could infinitesimally approach 0 (e.g. a great-horned owl at the bottom  
213 of the Mediterranean sea), a species would never have a score of 0 because at least *some* components  
214 of its niche are present: water, carbon, oxygen, etc. This presents an important consideration: just

215 because a species has an econative score greater than 0 at a location does not suggest that a species  
216 “belongs” there in any way—econative is a relative metric.

217 The econative concept is compatible with other definitions of native, new and old. I’ve already  
218 elaborated the nested relationship between classical native and econative, where classical nativeness  
219 refers to a temporal snapshot of econativeness. “Neonative” can also be viewed as a special case  
220 along the econative gradient, where a species becomes much more econative to an area and establishes  
221 populations. Essl et al. recognized the underlying continuum and argued that the categorization of  
222 nonnative, native, and neonative were motivated by utility [Essl et al., 2020]. The econative concept  
223 is consistent with this and simply makes explicit the underlying continuum that is being discretized.  
224 An alternate definition proposed by Gilroy et al. states that “native” and “nonnative” hinge on  
225 human-mediated transport. While econativeness does not directly consider method of transport, the  
226 incorporation of evolutionary processes and biotic composition into the econative niche concept should  
227 ensure that human-mediated transport across major biogeographic barriers would confer significantly  
228 lower econativeness.

229 In the following section, I’ll walk through three examples of the application of the econative con-  
230 cept which serve to clarify the concept, demonstrate its ethical relevance, and address some *prima*  
231 *facie* concerns about the implications for trans-continental translocations, species invasions, and the  
232 protection of incumbent species and ecosystems.

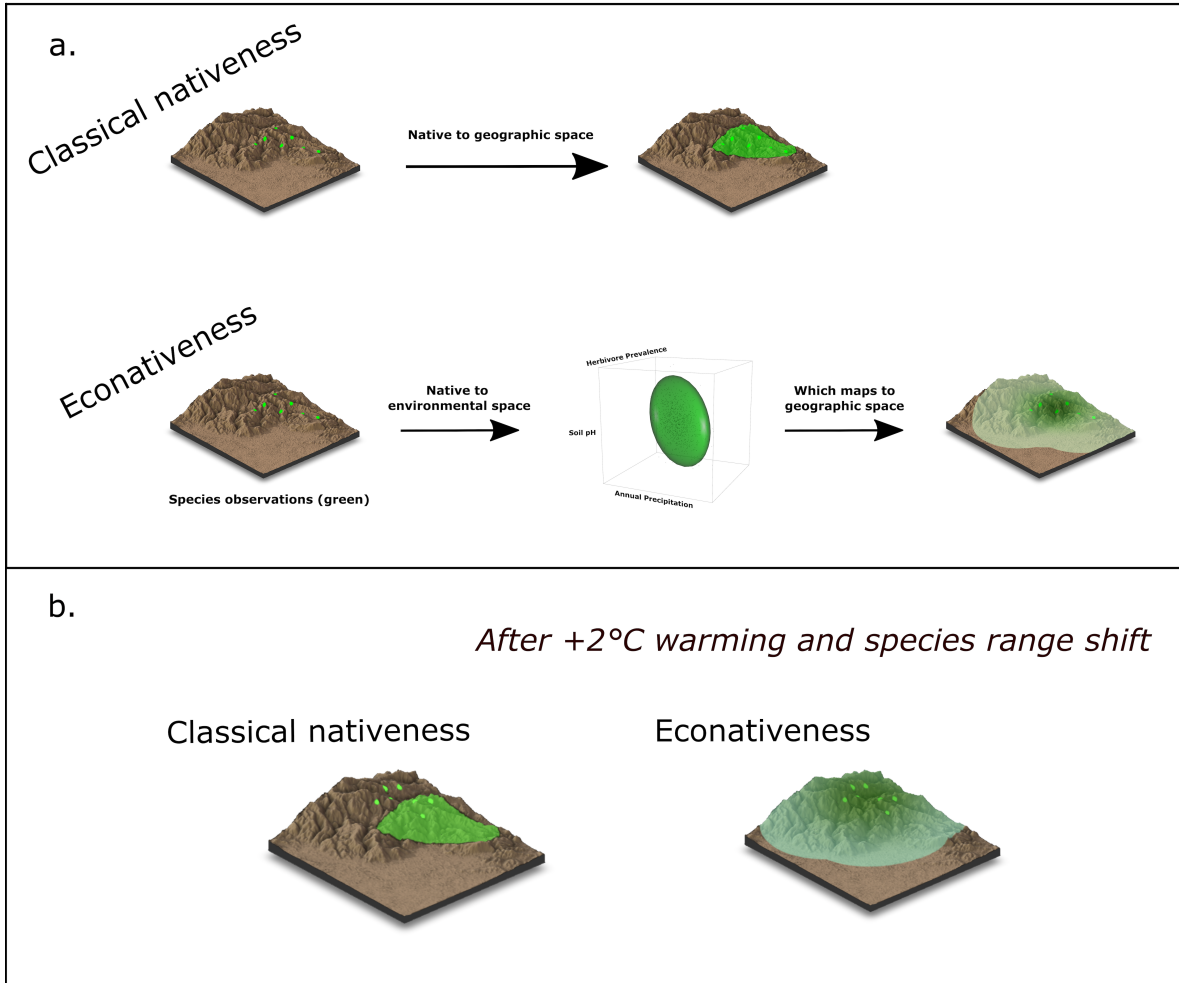


Figure 1: **Diagram comparing classical nativeness to econativeness.** a. Both accounts begin with observed species occurrences on a landscape (green dots, left), but diverge in the way that these occurrences are used to define a “native” range. In the classical understanding of nativeness, the native range is generally understood to be a contiguous geographic area which includes species presences and excludes species absences (top). Under econativeness (bottom), the species observations correspond to n-dimensional environmental space (the Hutchinsonian niche) to which the species is native. The native niche can then be mapped to geographic space, where all locations are on a continuum from less econative to more econative depending on how similar the environmental conditions are to the Hutchinsonian niche of the species. b. The different nativeness concepts yield significantly different results after a theoretical 2°C of climate warming and a resulting shift in the species distribution. The classical native concept cannot account for the individuals that migrated outside of their historical range and the significance of the lower-elevation, extirpated portion of the range is ambiguous.

233 **3 Applying “Econative”**

234 This reconsideration of nativeness has important implications for the attribution of value to species.  
 235 A few different types of value attributed to species are at least partly dependent on nativeness, and  
 236 these include instrumental values, natural historical values, and integral values (Table 1). Instrumental  
 237 values are those attributed to a species when it provides a service, natural historical value is derived  
 238 from an appreciation of the complex ecological relationships and evolutionary processes of a species  
 239 [Katz, 1997], and integral value applies when people have a preference for a species continued existence  
 240 in a way that is consistent with a person’s culture or worldview [Sandler, 2012].

<b>Intrinsic Value</b> <i>if valuable for its own sake</i>				<b>Instrumental Value</b> <i>if a means to an end that is intrinsically valuable</i>	
<b>Objective Intrinsic Value</b> <i>if valuable for its own sake regardless of someone's opinion</i>		<b>Subjective Intrinsic Value</b> <i>if valuable for its own sake based on someone's opinion</i>		<b>Present Value</b> <i>if currently a means to an end that is intrinsically valuable</i>	<b>Option Value</b> <i>if, at some point, potentially a means to an end that is intrinsically valuable</i>
<b>Natural Historical Value</b> <i>if a unique product of ecological relationships and evolutionary forces over long periods of time</i>	<b>Inherent Worth</b> <i>if possesses interests of its own that ought to be valued</i>	<b>Existence Value</b> <i>if someone has a preference for its continued existence</i>	<b>Integral Value</b> <i>if someone has a preference for its continued existence for reasons that are consistent with their culture or worldview</i>		

Table 1: **Nested table of values that may be attributed to species.** A commonly adopted framework for describing the types of values that might be attributed to species nested within the two broadest categories: intrinsic and instrumental. The necessary conditions for which a species is attributed each value is briefly summarized. This is largely based on the work of Ronald Sandler [Sandler, 2012].

241 Can these values still be attributed in-full to a species if classical nativeness is supplanted with  
 242 econative? The extremes of econativeness clearly inherit the same assortment of value from classical  
 243 nativeness because they are equivalent to the native/non-native binary. A species that is entirely  
 244 econative to a site possesses all of the native-dependent values that a native species would be attributed,  
 245 because under these conditions native and econative mean the same thing. Conversely, a species with  
 246 very little econativeness to a location has the same native-dependent value as a non-native species. The  
 247 main difference, and the source of its moral utility, is that econativeness allows for some fraction of these  
 248 values to be attributed to species in the increasingly frequent scenarios in which species are not fully  
 249 native or non-native. Imagine a montane subalpine forest community where, due to climate change, all  
 250 but a few species shift up-slope by 500m to a location where they have never historically co-occurred  
 251 before. ENM would yield an econative value close to 1—the abiotic and biotic environments are very  
 252 similar, but not equivalent. These shifted species are not 100% econative, and not native under the  
 253 classical understanding, but certainly some measure of natural historical value and the integral values  
 254 that arise from ecological and evolutionary relationships between the birds, spruces, soil microbes,  
 255 etc. are maintained. Not only is econativeness morally useful in the same conditions that classical  
 256 nativeness is, but also to a whole slew of scenarios that recent global change introduces. Using three  
 257 examples, I will continue to explore the utility of econativeness for species value attribution and the  
 258 conditions in which it outperforms the classical native concept.

259 The most straightforward example is one in which the environment has not changed much over the  
 260 last century: the species composition has remained stable and the climate, soil, disturbance regimes,  
 261 etc. are for the most part within the historical variance of the system over the last few hundred  
 262 years. Imagine that the only difference is that the mean annual temperature has increased by 0.2°C.  
 263 Would it make sense to introduce (or permit the introduction of) an endangered novel species to this  
 264 site? A comprehensive assessment of the econativeness of all species at the recipient site and the  
 265 introduced species yields an unsurprising conclusion: probably not. Although the increase in 0.2°C



266 may be outside the temperature range that some of the local species are adapted to—resulting in a  
267 decrease in econativeness for these species—most other dimensions of econativeness are intact. And  
268 even if the increase in 0.2°C makes the introduced species slightly more econative to the recipient  
269 site, it doesn’t become more econative than the incumbent species. Here, econativeness correlates  
270 with natural historical value, and other values attributed to the species that are dependent on its  
271 ecological and evolutionary situatedness. Econativeness is also correlated with the endangered species’  
272 likelihood of establishing—if the environment is very different from the conditions in which it evolved,  
273 then it is likely that it will be more difficult to survive and reproduce. The species composition  
274 component of econativeness, coupled with an analysis of the traits of the endangered species, might be  
275 useful for estimating its potential invasibility, and therefore ecological harm, in the recipient system.  
276 When isolated, the utility of the components of econativeness are obvious and have been used to  
277 estimate invasibility, habitat suitability, and value before—its strength and novelty is in its use as a  
278 unified concept. This thought experiment demonstrates that the econative concept strengthens the  
279 justification for protecting largely intact ecosystems and weakens justifications for the introduction or  
280 establishment of novel species.

281 Our next thought experiment concerns another extreme: an environment that is nearly unrecog-  
282 nizable from any historical state. An abandoned lot on the outskirts of Chicago may have few, if any,  
283 of the species that occurred there hundreds of years ago. Furthermore, the soil has contaminants from  
284 industrial activity, the winter temperature has increased by 1°C, and extreme precipitation events  
285 have increased by 40% [Illinois State Climatologist, 2021]. What species assemblage maximizes the  
286 value at this location? Econativeness alone is not sufficient for a complete value analysis because  
287 some values are largely independent from it, like many instrumental or aesthetic values. Perhaps a  
288 garden would provide the most value at this site. But an evaluation of econativeness is necessary for  
289 a complete value assessment, and a logical place to start would be to determine the econativeness of  
290 the present species, the historical species, and the historical species from nearby regions. The common  
291 “weed” species likely present at the lot, like garlic mustard (*Alliaria petiolata*) or cheatgrass (*Bromus*  
292 *tectorum*) have some measure of econativeness by simple virtue of being able to grow there—the cli-  
293 mate and soil conditions must be similar to the environments in which they evolved—but the biotic  
294 components (e.g. species interactions, co-occurrences) of their econativeness is quite low. It’s worth  
295 acknowledging evolution here, and the possibility that the Eurasian plants have started to form ecolog-  
296 ical and evolutionary relationships with local plants, animals, and abiotic features and that this may  
297 confer some additional measure of econativeness. Though this is probably minimal in so short a time.  
298 The historical species likely have historically co-occurring species more geographically close (perhaps  
299 somewhere else in Illinois) than the historically co-occurring partners of cheatgrass or garlic mustard  
300 (their historically co-occurring species are from Eurasia), and so have greater biotic econativeness.  
301 Consequently, those historical species that can survive in the parking lot conditions likely have more  
302 total econativeness. But due to ongoing climate change, it may be the case that the climate may  
303 be more similar to the niche of species from warmer and wetter parts of the midwestern U.S. that  
304 haven’t historically occurred near Chicago. These species, like the historically occurring ones, also  
305 have members of their historical biotic assemblages closer than Eurasia. It’s reasonable to think that  
306 at least some non-historical species that are newly suitable to the habitat due to climate change have  
307 greater econativeness to the abandoned lot than either the historical species or the present (“weedy”)  
308 species.

309 This thought experiment emphasizes how econativeness can implicitly account for the relevance  
310 of geographic distance between a species’ historical and introduced range. Eurasia is very distant  
311 from Illinois, both in geographic space and environmental space. Even between those areas where  
312 the climates are analogous, the species assemblages, geology, and ecological relationships might differ  
313 significantly. If the assemblage of species that cheatgrass co-occurs with in its historical range (e.g.  
314 the tens of thousands of plants, animals, and fungi around the Mediterranean sea) was translocated  
315 to central North America, then, perhaps, might cheatgrass become more econative to the Chicago lot  
316 than an Illinois species.

317 Our 3rd example concerns areas that are on the edge of species distributions actively shifting  
318 in response to climate change. Large swaths of conifer forests in the Sierra Nevada mountains of  
319 California are outside of the climate to which they’re adapted and conifers are being replaced by  
320 oaks and chaparral at the lower-elevation and warmer edge of their distribution [Hill et al., 2023]. In  
321 these areas of active and expected transitions, which species have the greatest value conferred by

322 ecological and evolutionary situatedness? The dichotomous understanding of nativeness is unhelpful  
323 here, and the classification of the migrating species as non-native underestimates their value. Essl  
324 et al. would call these migrating trees “neonative” [Essl et al., 2019], but it’s unclear what ethical  
325 significance neonative species have, particularly in comparison to the historical species that are being  
326 actively replaced. This example especially benefits from a gradient understanding of nativeness because  
327 in many of these sites, the difference in econativeness between historical and migrating species are  
328 rapidly shrinking. As the climate continues to change and the values derived from econativeness  
329 become more equivalent between the groups of species, the other values, not dependent on ecological  
330 or evolutionary situatedness, become more relevant to the land management decisions. The species  
331 that are attributed aesthetic, instrumental, or cultural value—independent of econativeness—may be  
332 an important foundation on which to build an understanding of the species that “belong” at these  
333 locations. Because econativeness is sensitive to biotic composition, the decision to prioritize particular  
334 species affects the econativeness of others. For example, if Jeffrey Pine (*Pinus jeffreyi*) is attributed a  
335 great deal of instrumental and integral value due to the vanilla-like perfume of its furrowed bark and  
336 is decidedly worth protecting in a particular area, then a number of species that historically co-exist  
337 with Jeffrey Pine are more econative to that area by virtue of Jeffrey Pine being present even if other  
338 environmental conditions have changed. This example is relevant to all areas undergoing vegetation  
339 transitions in response to climate change—at some point the econativeness of range-expanding species  
340 may approach equivalence to that of historical species, and the decision to slow, facilitate, or passively  
341 observe will hinge on values independent of nativeness.

## 342 4 Additional considerations

343 One important consideration is that the econative concept is more useful to conservation efforts that  
344 have greater emphasis on native-dependent values. Some conservation projects may not consider  
345 nativeness at all. Operating at an extreme of the “biodiversity-first” framework, nativeness could be  
346 irrelevant to the goal of maximizing the species count or genetic diversity at a locality (though it’s  
347 worth noting this is not a popular conservation goal). In a framework emphasizing naturalness above  
348 all else, the econative concept is only useful when considering species that are translocated without  
349 human intervention—if even then. However, conservation projects often have a diversity of goals and  
350 underlying values, and usually at least some of these values are dependent on nativeness.

351 Another important consideration is the proliferation of non-analog climate conditions and eco-  
352 logical communities expected in the coming years [Petrie et al., 2020, Williams and Jackson, 2007].  
353 Non-analog environmental conditions introduce serious challenges for ENM, primarily because the  
354 often-used correlative modeling framework has poor performance when models are extrapolated be-  
355 yond the domain of their training data [Fitzpatrick and Hargrove, 2009]. The practical application  
356 of econativeness is reduced when ENMs perform poorly. Realizing the impact of non-analog environ-  
357 ments on ENMs did not staunch the explosion of extrapolative ENMs in the decades since, but it did  
358 introduce important considerations of the limitations. Like extrapolative ENMs, the econative concept  
359 can still be informative while acknowledging the proper limitations and uncertainties introduced by  
360 novel environmental conditions. And as is also done with ENM, econativeness can be evaluated by its  
361 component dimensions (rainfall, predator occurrence, soil type), at least some of which will remain in  
362 analogous conditions.

## 363 5 Consequences for Assisted Migration

364 The econative concept has important implications for the debate on Assisted Migration (AM), the  
365 conservation-motivated movement of species to areas beyond their historical range. Of the many terms  
366 used to refer to this process (e.g. assisted colonization, managed relocation, etc.), Hällfors et al. argue  
367 that assisted migration (AM) is best when referring to the practice of “safeguarding biological diversity  
368 through the translocation of representatives of a species or population harmed by climate change to an  
369 area outside the indigenous range of that unit where it would be predicted to move as climate changes,  
370 were it not for anthropogenic dispersal barriers or lack of time” [Hällfors et al., 2014]. Despite being  
371 extremely controversial since its inception in 1985 [Peters and Darling, 1985] and “ignit[ing] long-  
372 smoldering tensions in American natural resources policy” [Camacho, 2010a], AM has been recognized

373 as a conservation tool by the International Union for Conservation of Nature [IUCN, 2013] and has  
374 been put into practice in a few isolated instances around the world (e.g. projects with *Torreya taxifolia*  
375 [Barlow, 2021] and *Pseudemydura umbrina* [Lewis, 2016]).

376 A number of ethical and ecological concerns have been raised in response, primarily regarding risk  
377 of ecological harm, practical efficacy, and the soundness of value-based justifications. Conservationists  
378 have good reason to be afraid of ecological harm: the movement of species outside of their historical  
379 range can lead to ecologically damaging invasions when the ecological conditions that kept a species  
380 population in control in its historical range (e.g. disease, predators, etc.) are not present in its intro-  
381 duced range. Populations under these conditions could increase dramatically and lead to significant  
382 damage in the recipient ecosystem [Courchamp et al., 2003]. Some argue that the risks of invasion  
383 far outweigh the potential benefit of AM [Ricciardi and Simberloff, 2009, Maier and Simberloff, 2016],  
384 and even proponents call for extensive risk assessment before implementation [Gallagher et al., 2015,  
385 Butt et al., 2021].

386 A number of ethical concerns compound the scientific and practical, but the primary debate centers  
387 on value: the potential benefits to target species, costs to recipient ecosystems, and underlying un-  
388 certainties [McLachlan et al., 2007, Schwartz et al., 2012]. In 2012, the Managed Relocation Working  
389 Group wrote that the first step towards developing an AM decision framework was the examination  
390 of the goals of conservation and their constituent values [Schwartz et al., 2012], and this effort is still  
391 ongoing. One important thread concerns which types of value, if any and to what extent, are main-  
392 tained during the process of AM [Schwartz et al., 2012, Sandler, 2012, Maier and Simberloff, 2016,  
393 Siipi and Ahteensuu, 2016]. This is central to the debate because an effective cost-benefit analysis—  
394 weighing the benefits of AM against the cost and risks of harm in the recipient ecosystem—is en-  
395 tirely dependent on an accurate analysis of value for the species and ecosystems involved. Indeed,  
396 three of the four dimensions of the AM evaluation tool proposed by the Managed Relocation Work-  
397 ing Group (focal impact, collateral impact, and acceptability) directly depend on the values at-  
398 tributed to the species involved [Richardson et al., 2009]. A chief concern by a number of ethicists  
399 is that many proponents of AM do not provide positive, value-based justification for its practice  
400 [Maier and Simberloff, 2016, Sandler, 2012]. For example, recent decision-making framework pub-  
401 lished by the U.S. National Park Service for the purposes of implementing AM includes a cost-benefit  
402 analysis equation where the “benefit” is exclusively a function of the reduction in the risk of extinction  
403 [Karasov-Olson et al., 2021]. Under these conditions, a zoo or conservatory might be the best choice  
404 for maximizing benefit (probability of species continuation) and minimizing the cost (ecosystem harm,  
405 resource use, etc.), but surely this is not an intended conclusion.

406 Proponents of AM don’t typically argue on behalf of the instrumental value of species [Lavrik, 2021]  
407 (the silviculture industry is a prominent exception [Thiffault et al., 2021, Gömöry et al., 2020]), so  
408 most of the debate centers on intrinsic value either implicitly or explicitly. Objective intrinsic value (by  
409 virtue of natural historical value) and integral values are the most relevant. Some argue (even a promi-  
410 nent critic like Sandler) that the aesthetic, cultural, spiritual, etc. properties composing the integral  
411 value of a species (a subjective intrinsic value) might be maintained by AM [Siipi and Ahteensuu, 2016,  
412 Sandler, 2010]. Natural historical value is more contentious—even in the historical range—but is also  
413 specifically referred to in some justifications of AM [Siipi and Ahteensuu, 2016]. Because so many  
414 proponents are not explicit about the value of species, we assume that they at least indirectly rely  
415 upon the natural historical value or integral value of species when not referring to instrumental value.

416 Critics argue that these values are either eroded or entirely lost by translocation outside of the his-  
417 torical range [Sandler, 2012, Maier and Simberloff, 2016] (although some interesting exceptions might  
418 include species with integral value to indigenous peoples that were translocated outside of their his-  
419 torical range, like taro in Hawaii or kiore in New Zealand). Both integral and natural historical value  
420 are described as being dependent on the species being *in situ* or in their native habitat because of  
421 the ecological and evolutionary relationships therein. However, the rigid and implicit framing of these  
422 properties as dichotomous and geographically explicit is not biologically justified, and the wholesale  
423 loss of those values leads to an underestimation of the value maintained by AM in many cases.

424 Econativeness is helpful here, and through it we acknowledge that species can maintain at least a  
425 fraction of their eco-evolutionary situatedness and the value that it confers beyond the geographic  
426 boundaries of their historically native range. The correlation between econativeness and native-  
427 dependent value must surely be complex and nonlinear, but in general we should expect that when  
428 a species is more econative to a locality it also maintains a greater number of eco-evolutionary rela-

429 tionships. This helps add nuance to a number of ambiguities in AM, while simultaneously clarifying  
430 and strengthening the justifications against widely-condemned practices like transcontinental translo-  
431 cations.

432 First, the econative concept could almost never be used to justify the AM of a species across major  
433 biogeographic barriers to a locality that it has not historically occurred. For reasons described earlier,  
434 the econativeness would be quite low (even if it's still within the species fundamental niche), and the  
435 native dependent values would be minimal. When accounting for the total costs and benefits, it would  
436 be extremely unlikely that the benefits would tip the balance. This conclusion is reachable without the  
437 econative concept, but the reason why a species does not belong across major biogeographic barriers  
438 is better articulated in this framework.

439 In the increasingly frequent case of climate-driven ecosystem transition, as detailed in the earlier  
440 example of conifer forests in the Sierra Nevada, econative is a helpful tool in AM decision-making. As  
441 the environment changes and the econativeness of the incumbent and migrating species become com-  
442 mensurate, other values and conservation goals become more likely to tip the cost-benefit calculations.  
443 Under the econative concept, we don't *prima facie* reject AM on the grounds of nativeness because we  
444 can appreciate that the migrating oaks have a number of eco-evolutionary relationships in the locality.  
445 At the same time, the incumbent conifers have less and less. All other values held equal, the dramatic  
446 shift in relative econativeness between the species might eventually prompt AM, in order to maximize  
447 total value among species at the locality. This is easiest to imagine in a scenario where the incumbent  
448 conifer species are slowly dying, becoming more susceptible to catastrophic wildfire, and competitively  
449 excluding oak trees and others that may otherwise be able migrate unassisted.

450 The last example builds on the abandoned Chicago lot thought experiment discussed earlier. Most  
451 of the species in this lot evolved a continent away, and the sum econativeness of the ecological commu-  
452 nity is quite low. When calculating the sum econativeness before and after the potential introduction  
453 of a nearby AM target species, the low econativeness of the incumbent species would be dwarfed by the  
454 econativeness of a migrating species from further south in Missouri. According to native-dependent  
455 values, there would be strong incentive to consider AM. Dramatically altered sites like this are the  
456 most easily justified recipient locations for AM, and may be helpful for establishing populations along  
457 the climate change-induced migration trajectory of a species. However, feasibility is a significant con-  
458 sideration in examples like this (abandoned lots probably aren't suitable for most endangered species),  
459 and many other values are perhaps more important than those related to conservation at an urban  
460 location like this (such as those related to affordable housing).

## 461 6 Conclusion

462 The challenges faced by the conservation community are broad and broadening. I think that “econa-  
463 tive” may be an important part of the lexicon of the next conservation paradigm and a useful tool  
464 for thinking through some of the difficult decisions that rapid global change brings. Conservation is  
465 motivated by the protection of the value of species and ecosystems, and a careful and comprehensive  
466 assessment of the value of each species at a site is necessary. Much of a species' value is tied to its  
467 ecological and evolutionary relationships, and the classical conception of nativeness is too inflexible to  
468 be useful for assessing a species value in a rapidly changing world. The practice of Assisted Migration  
469 has been so contentious because it forces some of the most difficult conservation questions to the fore.  
470 This proposed re-conceptualization of nativeness, ‘econativeness’, provides a nuanced and theoretically  
471 quantifiable framework that might help to think through these challenges. Important future directions  
472 include a coherent integration of evolution into the ecological niche concept, more complete ecological  
473 niche modeling of all the environmental components that describe a species niche, and the experimen-  
474 tal application of the econative concept to real-world conservation decisions. Ultimately, this work  
475 calls for—and contributes to—a careful consideration of which species belong where, and our role in  
476 stewarding these impending transitions.

477 **Supplementary information**

478 Not applicable

479 **Declarations**

480 The authors have no relevant financial or non-financial interests to disclose.

## References

- [Aaby and Ramsey, 2022] Aaby, B. H. and Ramsey, G. (2022). Three Kinds of Niche Construction. *The British Journal for the Philosophy of Science*, 73(2):351–372.
- [Barlow, 2021] Barlow, C. (2021). *Torreya* Guardians.
- [Barnosky et al., 2012] Barnosky, A. D., Hadly, E. A., Bascompte, J., Berlow, E. L., Brown, J. H., Fortelius, M., Getz, W. M., Harte, J., Hastings, A., Marquet, P. A., Martinez, N. D., Moers, A., Roopnarine, P., Vermeij, G., Williams, J. W., Gillespie, R., Kitzes, J., Marshall, C., Matzke, N., Mindell, D. P., Revilla, E., and Smith, A. B. (2012). Approaching a state shift in Earth’s biosphere. *Nature*, 486(7401):52–58.
- [Bentz et al., 2010] Bentz, B. J., Régnière, J., Fettig, C. J., Hansen, E. M., Hayes, J. L., Hicke, J. A., Kelsey, R. G., Negrón, J. F., and Seybold, S. J. (2010). Climate Change and Bark Beetles of the Western United States and Canada: Direct and Indirect Effects. *BioScience*, 60(8):602–613.
- [Bush et al., 2016] Bush, A., Mokany, K., Catullo, R., Hoffmann, A., Kellermann, V., Sgrò, C., McEvey, S., and Ferrier, S. (2016). Incorporating evolutionary adaptation in species distribution modelling reduces projected vulnerability to climate change. *Ecology Letters*, 19(12):1468–1478.
- [Butt et al., 2021] Butt, N., Chauvenet, A. L., Adams, V. M., Beger, M., Gallagher, R. V., Shanahan, D. F., Ward, M., Watson, J. E., and Possingham, H. P. (2021). Importance of species translocations under rapid climate change. *Conservation Biology*, 35(3):775–783.
- [Calflora, 2022] Calflora (2022). Information on California plants for education, research and conservation.
- [Camacho, 2010a] Camacho, A. E. (2010a). Assisted Migration: Redefining Nature and Natural Resource Law Under Climate Change. *Yale Journal on Regulation*, 27:87.
- [Camacho, 2010b] Camacho, A. E. (2010b). Transforming the Means and Ends of Natural Resources Management Adaptation and Resiliency in Legal Systems. *North Carolina Law Review*, 89(5):1405–1454.
- [Camacho et al., 2010] Camacho, A. E., Doremus, H., Mclachlan, J. S., and Minter, B. A. (2010). Reassessing Conservation Goals in a Changing Climate. *Issues in Science and Technology*, 26(4):21–26.
- [Courchamp et al., 2003] Courchamp, F., Chapuis, J.-L., and Pascal, M. (2003). Mammal invaders on islands: impact, control and control impact. *Biological Reviews*, 78(3):347–383.
- [Crees and Turvey, 2015] Crees, J. J. and Turvey, S. T. (2015). What constitutes a ‘native’ species? Insights from the Quaternary faunal record. *Biological Conservation*, 186:143–148.
- [Dodd and DeSilva, 2016] Dodd, R. S. and DeSilva, R. (2016). Long-term demographic decline and late glacial divergence in a Californian paleoendemic: *Sequoiadendron giganteum* (giant sequoia). *Ecology and Evolution*, 6(10):3342–3355.
- [Elith and Leathwick, 2009] Elith, J. and Leathwick, J. R. (2009). Species Distribution Models: Ecological Explanation and Prediction Across Space and Time. *Annual Review of Ecology, Evolution, and Systematics*, 40(1):677–697.
- [Essl et al., 2019] Essl, F., Dullinger, S., Genovesi, P., Hulme, P. E., Jeschke, J. M., Katsanevakis, S., Kühn, I., Lenzner, B., Pauchard, A., Pyšek, P., Rabitsch, W., Richardson, D. M., Seebens, H., van Kleunen, M., van der Putten, W. H., Vilà, M., and Bacher, S. (2019). A Conceptual Framework for Range-Expanding Species that Track Human-Induced Environmental Change. *BioScience*, 69(11):908–919.
- [Essl et al., 2020] Essl, F., Dullinger, S., Genovesi, P., Hulme, P. E., Jeschke, J. M., Katsanevakis, S., Kühn, I., Lenzner, B., Pauchard, A., Pyšek, P., Rabitsch, W., Richardson, D. M., Seebens, H., van Kleunen, M., van der Putten, W. H., Vilà, M., and Bacher, S. (2020). Distinct Biogeographic Phenomena Require a Specific Terminology: A Reply to Wilson and Sagoff. *BioScience*, 70(2):112–114.

- 529 [Fitzpatrick and Hargrove, 2009] Fitzpatrick, M. C. and Hargrove, W. W. (2009). The projection of  
530 species distribution models and the problem of non-analog climate. *Biodiversity and Conservation*,  
531 18(8):2255–2261.
- 532 [Gallagher et al., 2015] Gallagher, R. V., Makinson, R. O., Hogbin, P. M., and Hancock, N. (2015).  
533 Assisted colonization as a climate change adaptation tool: Assisted colonization and climate change.  
534 *Austral Ecology*, 40(1):12–20.
- 535 [Gömöry et al., 2020] Gömöry, D., Krajmerová, D., Hrivnák, M., and Longauer, R. (2020). Assisted  
536 migration vs. close-to-nature forestry: what are the prospects for tree populations under climate  
537 change? *Central European Forestry Journal*, 66(2):63–70.
- 538 [Hill and Hadly, 2018] Hill, A. P. and Hadly, E. A. (2018). Rethinking “Native” in the Anthropocene.  
539 *Frontiers in Earth Science*, 6:96.
- 540 [Hill et al., 2023] Hill, A. P., Nolan, C. J., Hemes, K. S., Cambron, T. W., and Field, C. B. (2023).  
541 Low-elevation conifers in California’s Sierra Nevada are out of equilibrium with climate. *PNAS*  
542 *Nexus*, 2(2):pgad004.
- 543 [Holt, 2009] Holt, R. D. (2009). Bringing the Hutchinsonian Niche into the 21st Century: Ecological  
544 and Evolutionary Perspectives. *Proceedings of the National Academy of Sciences of the United States*  
545 *of America*, 106:19659–19665.
- 546 [Hooke and Martín-Duque, 2012] Hooke, R. L. and Martín-Duque, J. F. (2012). Land transformation  
547 by humans: A review. *GSA Today*, 12(12):4–10.
- 548 [Hutchinson, 1957] Hutchinson, G. E. (1957). Concluding Remarks. *Cold Spring Harbor Symposia on*  
549 *Quantitative Biology*, 22(0):415–427.
- 550 [Hällfors et al., 2014] Hällfors, M. H., Vaara, E. M., Hyvärinen, M., Oksanen, M., Schulman, L. E.,  
551 Siipi, H., and Lehvävirta, S. (2014). Coming to Terms with the Concept of Moving Species Threat-  
552 ened by Climate Change – A Systematic Review of the Terminology and Definitions. *PLoS ONE*,  
553 9(7):e102979.
- 554 [Illinois State Climatologist, 2021] Illinois State Climatologist (2021). Climate Change in Illinois.
- 555 [IUCN, 2013] IUCN (2013). Guidelines for Reintroductions and Other Conservation Translocations.
- 556 [Jachowski et al., 2015] Jachowski, D. S., Kesler, D. C., Steen, D. A., and Walters, J. R. (2015).  
557 Redefining baselines in endangered species recovery: Endangered Species Baselines. *The Journal of*  
558 *Wildlife Management*, 79(1):3–9.
- 559 [Karasov-Olson et al., 2021] Karasov-Olson, A., Schwartz, M., Olden, J., Skikne, S., Hellmann, J.,  
560 Allen, S., Brigham, C., Buttke, D., Lawrence, D., Miller-Rushing, A., Morissette, J., Schuurman, G.,  
561 Trammell, M., and Hawkins-Hoffman, C. (2021). Ecological risk assessment of managed relocation  
562 as a climate change adaptation strategy. Technical report, National Park Service.
- 563 [Katz, 1997] Katz, E. (1997). *Nature as subject: human obligation and natural community*. Studies in  
564 social, political, and legal philosophy. Rowman & Littlefield, Lanham.
- 565 [Lavrik, 2021] Lavrik, M. (2021). Constructing regulation on assisted migration: findings from science  
566 and ethics. *SN Social Sciences*, 1(9):242.
- 567 [Lewis, 2016] Lewis, D. (2016). Relocating Australian tortoise sets controversial precedent. *Science*.
- 568 [Lowe, 2014] Lowe, G. D. (2014). Geologic history of the giant sequoia. *North America Research*  
569 *Group (Paleontology)*.
- 570 [Maier and Simberloff, 2016] Maier, D. S. and Simberloff, D. (2016). Assisted Migration in Normative  
571 and Scientific Context. *Journal of Agricultural and Environmental Ethics*, 29(5):857–882.
- 572 [McLachlan et al., 2007] McLachlan, J. S., Hellmann, J. J., and Schwartz, M. W. (2007). A Framework  
573 for Debate of Assisted Migration in an Era of Climate Change. *Conservation Biology*, 21(2):297–302.

- 574 [Morrow, 2024] Morrow, K. H. (2024). Niches and Niche Models. *The British Journal for the Philos-*  
575 *ophy of Science*.
- 576 [Nackley et al., 2017] Nackley, L. L., West, A. G., Skowno, A. L., and Bond, W. J. (2017). The  
577 Nebulous Ecology of Native Invasions. *Trends in Ecology & Evolution*, 32(11):814–824.
- 578 [Pereyra, 2020] Pereyra, P. J. (2020). Rethinking the native range concept. *Conservation Biology*,  
579 34(2):373–377.
- 580 [Peters and Darling, 1985] Peters, R. L. and Darling, J. (1985). The Greenhouse Effect and Nature  
581 Reserves. *BioScience*, 35(11):707–717.
- 582 [Petrie et al., 2020] Petrie, M. D., Bradford, J. B., Lauenroth, W. K., Schlaepfer, D. R., Andrews,  
583 C. M., and Bell, D. M. (2020). Non-analog increases to air, surface, and belowground temperature  
584 extreme events due to climate change. *Climatic Change*, 163(4):2233–2256.
- 585 [Reyer et al., 2015] Reyer, C. P. O., Brouwers, N., Rammig, A., Brook, B. W., Epila, J., Grant, R. F.,  
586 Holmgren, M., Langerwisch, F., Leuzinger, S., Lucht, W., Medlyn, B., Pfeifer, M., Steinkamp, J.,  
587 Vanderwel, M. C., Verbeeck, H., and Villeda, D. M. (2015). Forest resilience and tipping points at  
588 different spatio-temporal scales: approaches and challenges. *Journal of Ecology*, 103(1):5–15.
- 589 [Ricciardi and Simberloff, 2009] Ricciardi, A. and Simberloff, D. (2009). Assisted colonization is not  
590 a viable conservation strategy. *Trends in Ecology & Evolution*, 24(5):248–253.
- 591 [Richardson et al., 2009] Richardson, D. M., Hellmann, J. J., McLachlan, J. S., Sax, D. F., Schwartz,  
592 M. W., Gonzalez, P., Brennan, E. J., Camacho, A., Root, T. L., Sala, O. E., Schneider, S. H., Ashe,  
593 D. M., Clark, J. R., Early, R., Etterson, J. R., Fielder, E. D., Gill, J. L., Minter, B. A., Polasky, S.,  
594 Safford, H. D., Thompson, A. R., and Vellend, M. (2009). Multidimensional evaluation of managed  
595 relocation. *Proceedings of the National Academy of Sciences*, 106(24):9721–9724.
- 596 [Roy et al., 2023] Roy, H. E., Pauchard, A., Stoett, P., Renard Truong, T., Bacher, S., Galil, B. S.,  
597 Hulme, P. E., Ikeda, T., Sankaran, K., McGeoch, M. A., Meyerson, L. A., Nuñez, M. A., Ordonez,  
598 A., Rahlao, S. J., Schwindt, E., Seebens, H., Sheppard, A. W., and Vandvik, V. (2023). IPBES  
599 Invasive Alien Species Assessment: Summary for Policymakers. Technical report, Zenodo.
- 600 [Sandler, 2010] Sandler, R. (2010). The Value of Species and the Ethical Foundations of Assisted  
601 Colonization. *Conservation Biology*, 24(2):424–431.
- 602 [Sandler, 2012] Sandler, R. (2012). *The Ethics of Species: An Introduction*. Cambridge University  
603 Press, Cambridge.
- 604 [Scheffer et al., 2009] Scheffer, M., Bascompte, J., Brock, W. A., Brovkin, V., Carpenter, S. R., Dakos,  
605 V., Held, H., van Nes, E. H., Rietkerk, M., and Sugihara, G. (2009). Early-warning signals for critical  
606 transitions. *Nature*, 461(7260):53–59.
- 607 [Schwartz et al., 2012] Schwartz, M. W., Hellmann, J. J., McLachlan, J. M., Sax, D. F., Borevitz,  
608 J. O., Brennan, J., Camacho, A. E., Ceballos, G., Clark, J. R., Doremus, H., Early, R., Etterson,  
609 J. R., Fielder, D., Gill, J. L., Gonzalez, P., Green, N., Hannah, L., Jamieson, D. W., Javeline, D.,  
610 Minter, B. A., Odenbaugh, J., Polasky, S., Richardson, D. M., Root, T. L., Safford, H. D., Sala, O.,  
611 Schneider, S. H., Thompson, A. R., Williams, J. W., Vellend, M., Vitt, P., and Zellmer, S. (2012).  
612 Managed Relocation: Integrating the Scientific, Regulatory, and Ethical Challenges. *BioScience*,  
613 62(8):732–743.
- 614 [Shackelford et al., 2013] Shackelford, N., Renton, M., Perring, M. P., and Hobbs, R. J. (2013). Model-  
615 ing disturbance-based native invasive species control and its implications for management. *Ecological*  
616 *Applications*, 23(6):1331–1344.
- 617 [Siipi and Ahteensuu, 2016] Siipi, H. and Ahteensuu, M. (2016). Moral Relevance of Range and Nat-  
618 uralness in Assisted Migration. *Environmental Values*, 25(4):465–483.
- 619 [Soule, 1985] Soule, M. E. (1985). What Is Conservation Biology? *BioScience*, 35(11):9.



- 620 [Thiffault et al., 2021] Thiffault, N., Raymond, P., Lussier, J.-M., Aubin, I., Royer-Tardif, S.,  
621 D’Amato, A. W., Doyon, F., Lafleur, B., Perron, M., Bousquet, J., Isabel, N., Carles, S., Lupien,  
622 P., and Malenfant, A. (2021). Adaptive Silviculture for Climate Change: From Concepts to Reality  
623 Report on a symposium held at Carrefour Forêts 2019. *The Forestry Chronicle*, 97(01):13–27.
- 624 [Trappes, 2021] Trappes, R. (2021). Defining the niche for niche construction: evolutionary and eco-  
625 logical niches. *Biology & Philosophy*, 36(3):31.
- 626 [U.S. National Park Service, 2021] U.S. National Park Service (2021). Giant Sequoias and Fire.
- 627 [Walker et al., 2023] Walker, X. J., Okano, K., Berner, L. T., Massey, R., Goetz, S. J., Johnstone,  
628 J. F., and Mack, M. C. (2023). Shifts in Ecological Legacies Support Hysteresis of Stand Type  
629 Conversions in Boreal Forests. *Ecosystems*.
- 630 [Warren, 2007] Warren, C. R. (2007). Perspectives on the ‘alien’ versus ‘native’ species debate: a  
631 critique of concepts, language and practice. *Progress in Human Geography*, 31(4):427–446.
- 632 [Williams and Jackson, 2007] Williams, J. W. and Jackson, S. T. (2007). Novel climates, no-analog  
633 communities, and ecological surprises. *Frontiers in Ecology and the Environment*, 5(9):475–482.
- 634 [Wilson, 2020] Wilson, J. R. U. (2020). Definitions Can Confuse: Why the “Neonative” Neologism Is  
635 Bad for Conservation. *BioScience*, 70(2):110–111.
- 636 [Wisz et al., 2013] Wisz, M. S., Pottier, J., Kissling, W. D., Pellissier, L., Lenoir, J., Damgaard, C.,  
637 Dormann, C. F., Forchhammer, M. C., Grytnes, J.-A., Guisan, A., Heikkinen, R. K., Høye, T. T.,  
638 Kühn, I., Luoto, M., Maiorano, L., Nilsson, M.-C., Normand, S., Öckinger, E., Schmidt, N. M.,  
639 Mette Termansen, Termansen, M., Allan Timmermann, Timmermann, A., Wardle, D. A., Aastrup,  
640 P., and Svenning, J.-C. (2013). The role of biotic interactions in shaping distributions and realised  
641 assemblages of species: implications for species distribution modelling. *Biological Reviews*, 88(1):15–  
642 30.